

Pilot-scale mineralization of flue gas desulfurization gypsum in waste alkali medium

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Abstract

Flue gas desulfurization gypsum (FGDG) mainly comes from the FGD process in thermal power plants, smelters and so on. In this study, FGDG was used as raw material of CO₂ mineralization in NaOH medium in a pilot-scale bubble column reactor. The effects of pH value, CO₂ flow, reaction temperature, liquid level in bubble column reactor on the CO₂ mineralization process were investigated. The obtained crystal, morphology and particle size distribution of mineralization products were also studied. The results show that pH has a significant effect on CO₂ mineralization capacity. CO₂ flow and liquid have a slight effect on the particle size distribution of mineralized products. The crystal, morphology and particle size distribution of the obtained mineralization products were also observed. XRD and SEM show that the FGDG mineralization product is a single clustered crystal. The pilot results provide an optimized evidence for FGDG mineralization.

Keyword: Flue gas desulfurization gypsum (FGDG), NaOH, pilot-scale, mineralization.

1. Introduction

The use of fossil energy emits a large amount of greenhouse gases (CO₂, CH₄, N₂O, O₃, etc.) with the continuous development of global industrialization, which leads to increasing global warming and poses a significant threat to the ecological environment as well as the survival of human beings [1, 2], and thus the reduction of CO₂ emissions is receiving more and more attention [3, 4]. The fundamental way to reduce CO₂ emissions is to use clean energy instead of fossil fuels, however, the large-scale application of clean energy cannot be realized in a short time. Therefore CO₂ storage is a necessary means to achieve CO₂ emission reduction at this stage[5]. Mineral sequestration (mineralization) generates stable products, which can ensure permanent, safe and efficient CO₂ sequestration with sufficiently low risk of leakage, and has been widely focused on by many researchers[6], and this technology performs CO₂ mineralization by precipitating stable carbonate compounds, such as calcium carbonate and magnesium carbonate[7, 8]. Certain natural minerals and industrial solid

wastes containing large amounts of Ca²⁺ are used as mineralization feedstock. The environment required for CO₂ mineralization of natural products is harsh and mining, milling and separation make them costly[9]. In contrast, industrial solid wastes such as fly ash [10, 11], waste gypsum [12-14], steel slag [15], and calcium carbide slag [16] have been considered by researchers to have great potential for mineralization due to their relatively high reactivity and easy availability.

FGDG is a by-product of the flue gas wet desulfurization process in coal-fired power plants, and its main component is CaSO₄·2H₂O, which can be up to 93 wt%. At present, China's annual emissions of desulfurized gypsum are about 80 million tons, with a utilization rate of 78%, and a cumulative stockpile of about 100 million tons[17]. Scholars at home and abroad have conducted a lot of studies on the mineralization of FGDG for CO₂ sequestration. Lee [18] et al. used direct wet mineralization of FGDG to investigate the effects of process parameters such as ammonia concentration, CO₂ flow rate, solid-liquid ratio, and other process parameters on the

conversion rate and CO₂ mineralization efficiency of FGDG. Altiner [14] et al. adopted a "two-step method", that is, NaOH and FGDG were thoroughly mixed to form the Ca(OH)₂ suspension, and then CO₂ was introduced into the Ca(OH)₂ suspension to form CaCO₃ particles, but there were impurities in the product, which affected the purity of the product. Wing [12, 19] et al. proposed an ultrasonic technique to study the effects of different reaction media on the mineralization process and the crystalline shape of the product. Ding [20] studied the process flow of FGDG mineralization of CO₂ in NH₄Cl-H₂O system based on the theory of solubility equilibrium of salt solution and insoluble electrolyte, the carbonation rate can reach 98%, in which the carbonation reaction filtrate can be recycled. Wang [21] proposed a process of amine-promoted carbonation of waste gypsum in conjunction with the regeneration of amine by bipolar membrane electrodialysis (BMED), and seven different types of typical amines were selected as reaction media in the CO₂ mineralization. In spite of inspiring results obtained in terms of carbonation efficiency and CO₂ mineralization capacity, few studies on CO₂ mineralization at pilot scale have been unveiled so far. This paper aims to realize CO₂ mineralization at pilot scale in which FGDG as raw material and NaOH effluent from power plants as reaction medium.

2. Materials and methodology

2.1 Materials

The FGDG used in this study was produced from a thermal power plant in China, the FGDG simple was dried at 105 °C for 8 h to remove surface moisture and sieved to 60 mesh. The purity of CaSO₄ in FGDG used in the study is 97%, as determined by wet analysis. X-ray diffraction showed that the main component of the FGDG samples was calcium sulfate dihydrate, as shown in Fig.1. The particle size was measured by using a particle size analyzer (Bettersize 3000

Plus), the average size of the FGDG samples (d₅₀) was 31.18 μm, with 90% of the particles smaller than 83.98 μm. The composition of the sample was analyzed by XRF (ZSX Primuss II, Rigaku), and the FGDG impurities mainly contained SiO₂ and Al₂O₃, as shown in Table 1.

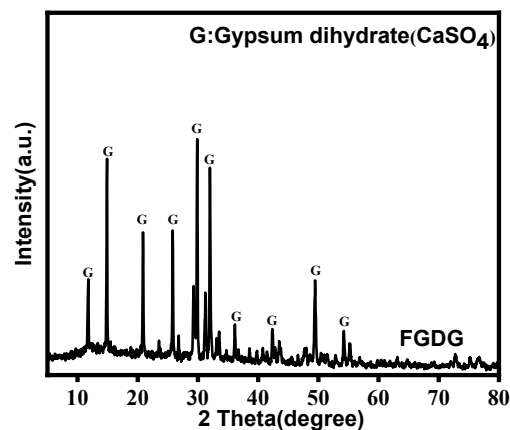


Fig.1 XRD pattern of FGDG.

2.2 CO₂ mineralization experiment

NaOH effluent from power plants pH 8, 10 and 12 were applied for the experiments respectively. The NaOH effluent was pumped into the pilot-scale reactor, as shown in Fig.2. After heating the NaOH effluent to the desired temperature, a weighed FGDG simple was added into the NaOH effluent through the auger, to prepare a slurry, and then 99% CO₂ gas was bubbled into the slurry with vigorous stirring. The effects of reaction temperature (40 °C, 60 °C, 80 °C), pH (8, 10, 12), CO₂ flow rate (20 L·h⁻¹, 40 L·h⁻¹, 60 L·h⁻¹) and liquid level (30 cm, 40 cm, 50 cm) on the FGDG mineralization were investigated. Detailed experimental conditions are shown in Table 2.

For example, pH8-T40-r20-h30 means that the experiments were carried out under the conditions of pH=8, reaction temperature 40 °C, CO₂ flow rate 20 L·h⁻¹ and liquid level 30 cm. Samples were taken from the exit at a given reaction time for determination (2, 4, 6, 15, 30, and 60 min) and pH was measured at reaction time points (2, 4, 6, 8, 10, 15, 20, 25, 30, 40, 50, and 60 min).

Table 1. Chemical composition of FGDG

Material	CaO	SO ₃	SiO ₂	Al ₂ O ₃	MgO	Na ₂ O	K ₂ O	LOI*
wt%	46.02	46.15	3.6	1.56	0.55	0.17	0.26	1.69

Table 2. Reaction condition in pilot scale experiments

Reaction condition	1	2	3
Temperature, T/°C	40	60	80
pH	8	10	12
CO ₂ flow, r/L·h ⁻¹	20	40	60
Liquid level, h/cm	30	40	50

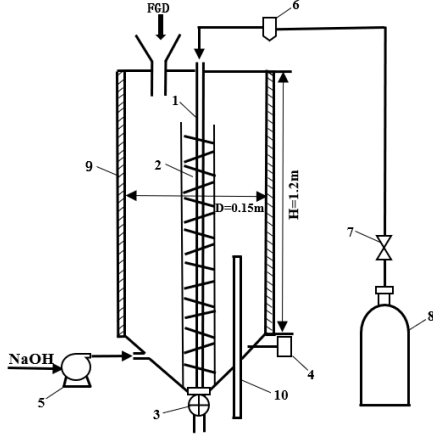


Fig.2 Structure of a pilot-scale mineralization reactor: 1- gas tubes, 2- heating coils, 3- valve, 4- pH meter, 5- liquid pump, 6- gas flow meter, 7- gas pressure reducing valves, 8- CO₂ cylinders, 9- insulation, 10- reclaimer tube

2.3 Characterization of product and calculation of storage capacity

All samples were separated from liquid by a high-speed centrifuge, and the obtained solid samples were dried at 180 °C for 8 h. the purity of CaCO₃ in the product was calculated from the weight loss of solid samples during the thermogravimetric process under N₂ atmosphere conditions. Pane and Hansen concluded that the decomposition of the CaCO₃ occurred at 600 °C to 780 °C[22]. The particle size analyzer (Bettersize3000Plus) was used to determine the particle size distribution of the mineralization products, and the morphology of the products was observed by scanning electron microscopy (SEM). The mass of CaCO₃ in the sample at time t is calculated by the following equation:

$$m_{t-\text{CaCO}_3} = \frac{\Delta m_{\text{CO}_2}}{M_{\text{CO}_2}} \times M_{\text{CaCO}_3} \quad (1)$$

Δm_{CO_2} is the weight of CO₂ that escapes when CaCO₃ decomposes during thermogravimetry, $m_{t-\text{CaCO}_3}$ is the mass of CaCO₃ in the

sample of products. M_{CaCO_3} is the relative molecular weight of CaCO₃. M_{CO_2} is the relative molecular weight of CO₂. The purity of CaCO₃ in the sample at time t is calculated by the following equation:

$$P_{t-\text{CaCO}_3} = \frac{m_{t-\text{CaCO}_3}}{m_p} \quad (2)$$

Here, $P_{t-\text{CaCO}_3}$ is the purity of CaCO₃ in the sample of products, m_p is the mass of sample of products.

$$\rho_t = \frac{P_{t-\text{CaCO}_3}}{P_{\text{C-CaCO}_3}} \quad (3)$$

ρ_t is the CO₂ mineralization efficiency of FGDG, $P_{\text{C-CaCO}_3}$ is the purity of CaCO₃ when FGDG was 100% mineralized. The purity of CaSO₄ in FGDG used in the study is 97%, as determined by wet analysis. Therefore, the purity of the reaction product should be 95.9% for 100% mineralization.

3 Results and Discussion

3.1 Effect of reaction parameters on CO₂ mineralization efficiency

3.1.1 Effect of initial pH on CO₂ mineralization efficiency

The effect of initial pH of NaOH effluent on FGDG mineralization was evaluated by varying initial pH which were 8, 10, 12 (pH8, pH10 and pH12).

Fig.3a shows the variation of slurry pH with time. As CO₂ was introduced into the slurry, the reaction began, and OH⁻ ions in the slurry was consumed, resulting in a decrease in the pH of the slurry. The pH of the slurry gradually stabilizes to about 7 after about 6 min of continuous reaction. The final pH of the slurry slightly increased with the initial pH of the NaOH effluent increased. The

CO₂ mineralization efficiency of FGDG was also affected by the initial pH of the NaOH effluent. After about 6 min of reaction, the mineralization reaction was basically completed, and then the reaction approached saturation state, as shown in Fig.3b.

The CO₂ mineralization efficiency of FGDG increased slightly with the continuous influx of CO₂ after 6 min of reaction. As the initial pH of NaOH effluent increased from 8 to 12, the CO₂ mineralization efficiency increased from 9.8% to 67.76%. The amount of FGDG involved in the mineralization reaction was small due to the low concentration of OH⁻ when the pH of the NaOH effluent was respectively 8 and 10. Under all conditions, the CO₂ mineralization efficiency of FGDG and slurry pH tended to stabilize at around 6 min, so this time point was considered the termination point of the basically completed mineralization reaction, as shown in Fig.3a.

The above results indicate that a high pH of NaOH effluent would be beneficial for the mineralization reaction to generate higher purity of CaCO₃ in the product, and the mineralization reaction would be more thorough. The high pH of NaOH effluent was ideal for obtaining higher CO₂ mineralization efficiency. The concentration of OH⁻ in the effluent was higher, the leaching Ca²⁺ was more, and the CO₂ mineralization efficiency was higher when the initial pH of NaOH effluent was 12 [23, 24].

3.1.2 Effect of CO₂ flow rate on CO₂ mineralization efficiency

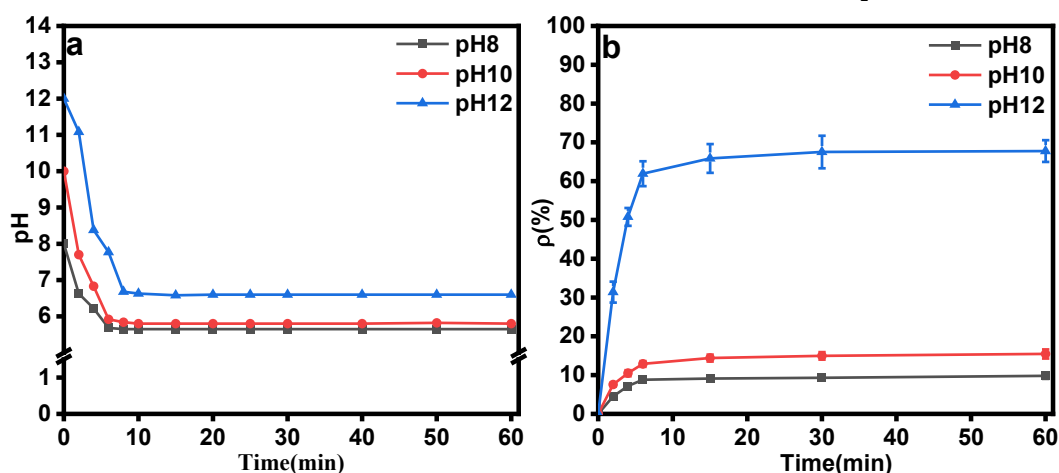


Fig.3 a: The pH value curves depended on time b: Effect of initial pH on the mineralization efficiency. (Conditions:40 °C temperature, 20 L·h⁻¹ CO₂ flow rate, 30 cm liquid level)

The effects of the CO₂ flow rate on CO₂ mineralization was evaluated by varying CO₂ flow rate which were 20, 40 and 60 L·h⁻¹ (r20, r40 and r60).

In all cases, the pH of the slurry decreased until it stabilized to 7 at around 6 min, as shown in Fig.4a. The CO₂ mineralization efficiency tended to stabilize when the reaction lasted for about 6 min. However, the final CO₂ mineralization efficiency decreased from 67.76% to 38.35% when the CO₂ flow rate was increased from 20 L·h⁻¹ to 60 L·h⁻¹, as shown in Fig.4b. Increasing CO₂ flow rate was not conducive to achieving high CO₂ mineralization efficiency of FGDG.

The diameter of the bubble increased and contact area between CO₂ and the slurry decreased with the increased of CO₂ flow rate [25], and the contact time between the CO₂ and the slurry decreases, resulting in the reduction of the mineralization efficiency in reaction products [26], which was not conducive to CO₂ mineralization. However, at 80 °C, the reaction temperature was too high, which reduced the solubility of CO₂ in the liquid, resulting in a decrease in carbonate ion concentration and a decreased in the CO₂ mineralization efficiency [25].

3.1.3 Effect of reaction temperature on the CO₂ mineralization efficiency

The effects of the reaction temperature of the NaOH effluent on the CO₂ mineralization was evaluated by varying the reaction temperature of the NaOH effluent. The CO₂ mineralization

efficiency tended to stabilize and the pH of slurry reached 7 after about 6 min when the reaction temperature is 40 °C and 60 °C, as shown in Fig.5a. The time that the CO₂ mineralization efficiency and the pH of slurry to reached stability increased to about 15 min at 80 °C. When the reaction temperature was increased from 40 °C to 60 °C, the CO₂ mineralization efficiency raised from 67.76% to 78.31%. While the reaction temperature was increased from 60 °C to 80 °C, the CO₂ mineralization efficiency was decreased from 75.5% to 55.62%, as shown in Fig.5b. This example demonstrates that increasing the reaction temperature would be beneficial for accelerating the mineralization reaction rate and high CO₂ mineralization efficiency. However, excessive reaction temperature was not conducive to CO₂ mineralization.

From the kinetic point of view, when the temperature was increased from 40 °C to 60 °C,

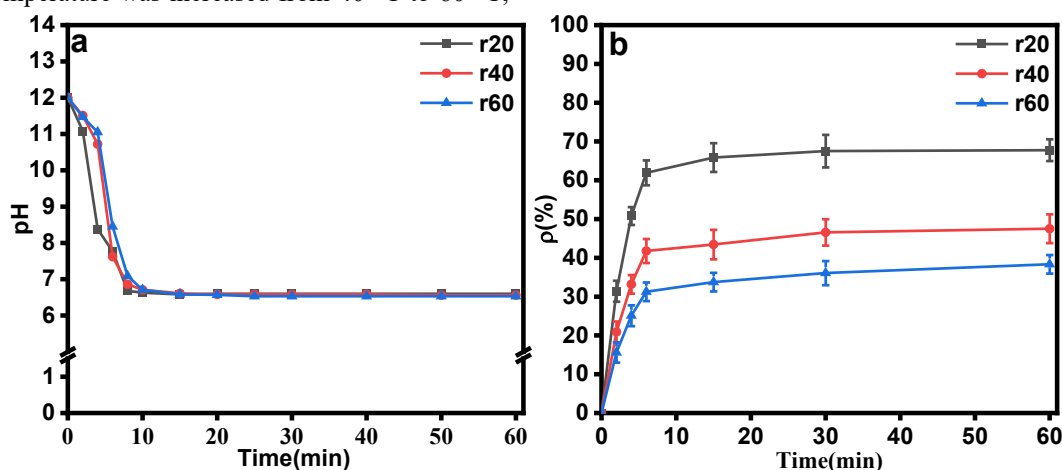


Fig.4 a: The pH value curves depended on time b: Effect of CO₂ flow rate on the mineralization efficiency. (Conditions:40°C temperature, pH:12,30 cm liquid level)

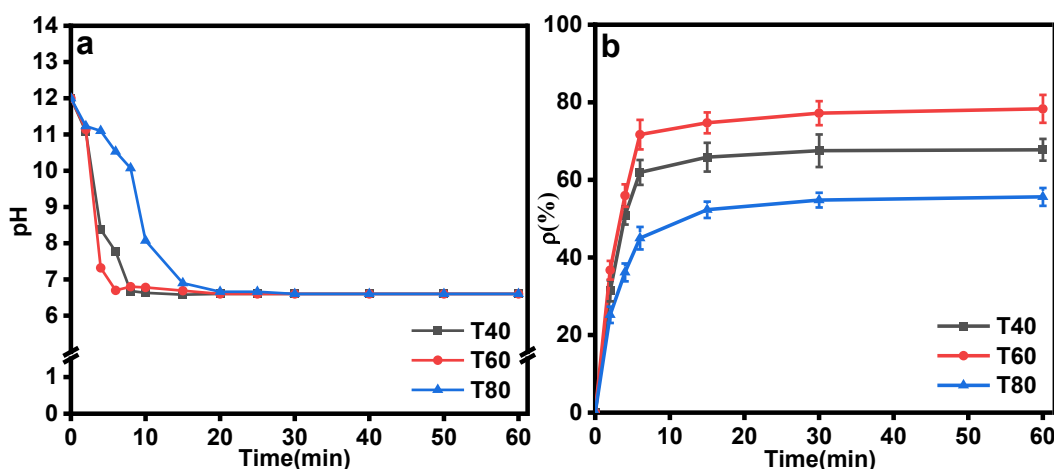


Fig.5 a: The pH value curves depended on time b: Effect of reaction temperature on the mineralization efficiency. (Conditions: 20 L·h⁻¹ CO₂ flow rate, pH:12, 30 cm liquid level)

the mass transfer rate could be increased, the thermal motion of the molecule can be promoted and the average kinetic energy of the molecule could be increased, so as to improve the CO₂ mineralization capacity and increased the reaction rate[27].

3.1.4 Effect of liquid level on the CO₂ mineralization efficiency

The effects of the liquid level of the NaOH effluent in the reactor on CO₂ mineralization was evaluated by varying liquid level which were 30, 40 and 50 cm. The difference of liquid level can lead to different capacities of NaOH effluent in the reactor during the experiment. Therefore, it is necessary to change the quantity of FGDC added to ensure that the solid-liquid ratio remains unchanged and the experiment proceeds smoothly. The change of liquid level had no significant

process. After about 6 min of reaction, the pH of the slurry decreased to 7, and the decline curve was consistent, as shown in Fig.6a. At the same time, the CO₂ mineralization efficiency tended to stabilize after about 6 min of reaction. When the liquid level increased from 30 to 50 cm, the CO₂ mineralization efficiency increased from 78.31% to 92.15%, as shown in Fig.6b. The above results indicate that a high liquid would be advantageous for high CO₂ mineralization efficiency. However high liquid level could not accelerate CO₂ mineralization reaction rate.

The contact time between the incoming CO₂ and the slurry in the reactor increased with the increased of liquid level, which increased the reaction time of the gas-liquid-solid three-phase in the reactor, thereby promoting the CO₂ mineralization efficiency [28].

3.2 Effect of reaction parameters on product particle size

3.2.1 Effect of initial pH on product particle size

The effect of the initial pH of NaOH effluent on the particle size distribution of the products was investigated by changing the initial pH (pH8, pH10 and pH12). Fig.7 shows the particle size distribution of the product which was generated under different initial pH conditions. The product was in a bimodal particle size distribution and the peaks value were about 2μm and 25μm. When the initial pH were 8, 10 and 12, the d₅₀ of the products was 13.4 μm, 11.7 μm and 10.4 μm,

respectively. The d₅₀ of the products decreased by 29.97% as the initial pH increased from 8 to 12. Therefore, the increased of initial pH would be favorable for obtaining small particle product.

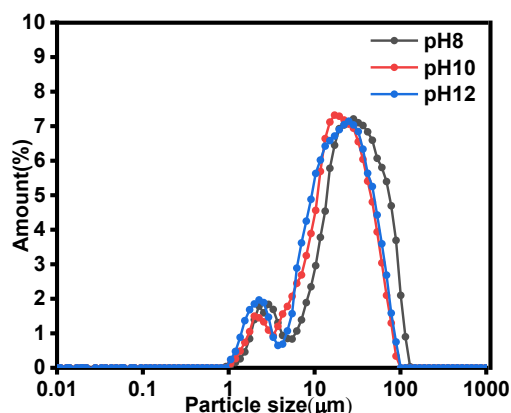


Fig.7 Effect of the pH on particle size distribution of products. (Conditions: 20 L·h⁻¹ CO₂ flow rate 40 °C temperature, 30 cm liquid level)

3.2.2 Effect of CO₂ flow rate on product particle size

The CO₂ flow rate also affected the particle size distribution of the products. Fig.8 shows the influence of CO₂ flow rate on particle size distribution of products. The bimodal peak values of the product were at 1.9 μm and 28.4 μm, and the d₅₀ value of product was 10.4 μm at 20 L·h⁻¹. The bimodal peak values of the product were about 3.4 μm and 32.2 μm, then the d₅₀ value of products were 17.1 μm and 19.4 μm when the CO₂ flow rate were 40 L·h⁻¹ and 60 L·h⁻¹, respectively. The bimodal values of product particle size shifted to the right with the increased

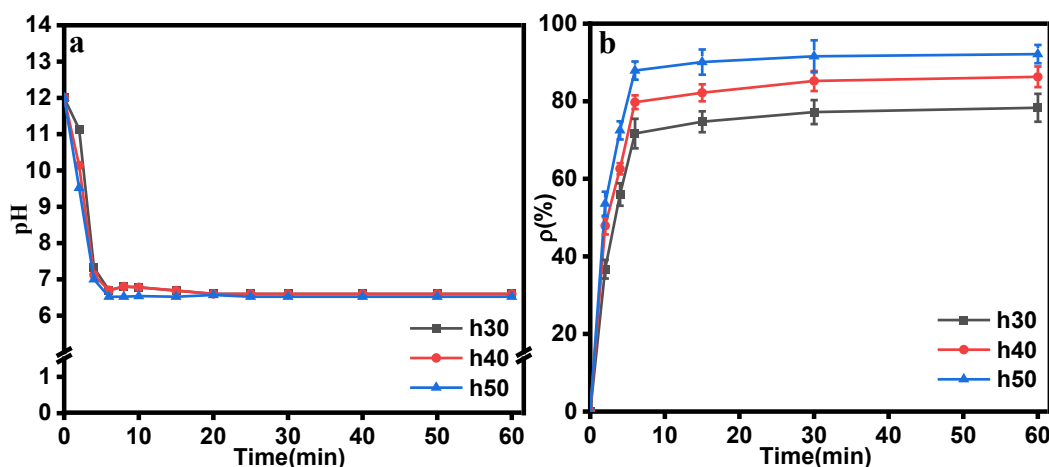


Fig.6 a: The pH value curves depended on time b: Effect of liquid level on the mineralization efficiency (Conditions: 20 L·h⁻¹ CO₂ flow rate, 60°C temperature, pH:12)

of CO₂ flow rate. The d₅₀ of the product decreased by 86.55% as the CO₂ flow rate increased from 20 L·h⁻¹ to 60 L·h⁻¹. The increased of CO₂ flow rate would be increased the percentage of large particles in the product, which was not conducive to the generation of small particles.

The increased of CO₂ flow rate accelerated crystal growth, so the low CO₂ flow rate was conducive to the acquisition of small particle product. The increased of CO₂ flow rate led to more intense churning and stirring of the slurry[29], but at the same time, the drag force of the floating would be greater. Particles with large mass and large particle size was easy to reach high liquid level positions, resulting in large particle size of the product.

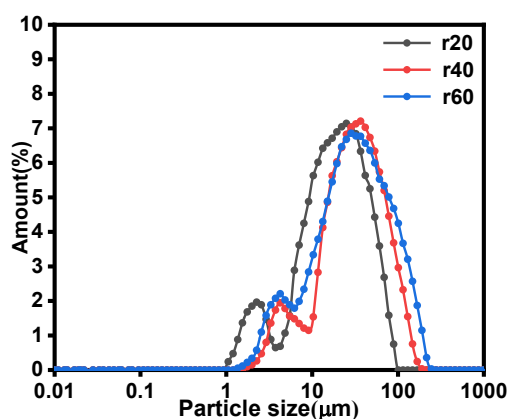


Fig.8 Effect of the CO₂ flow rate on particle size distribution. (Conditions:40°C temperature, pH:12,30 cm liquid level)

3.2.3 Effect of reaction temperature on product particle size

Fig.9 shows that the effect of reaction temperature on particle size distribution in the product. Under all conditions, the product exhibited a bimodal distribution, then the bimodal peak values ranging respectively from 2.3 μm to 2.9 μm and 22.5 μm to 28.4 μm. The d₅₀ value of products varied between 10.4 μm to 11.3 μm. The results indicated that the effects of reaction temperature on the particle size distribution of the product was not obvious[30].

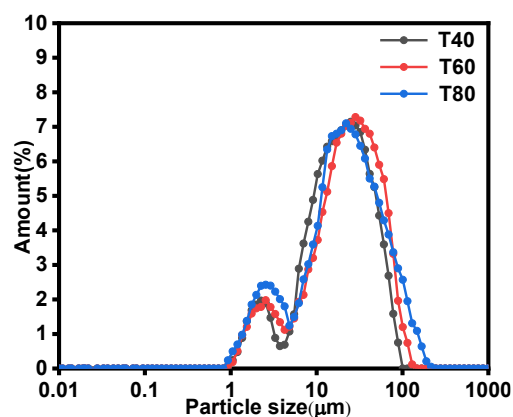


Fig.9 Effect of the reaction temperature on the particle size distribution. (Conditions: 20 L·h⁻¹ CO₂ flow rate, pH:12,30 cm liquid level)

3.2.4 Effect of liquid level on product particle size

In this stage, study on the influence of liquid level on product particle size distribution by changing the liquid level inside the reactor. Fig.10 shows a bimodal distribution of product under different conditions of liquid level. The bimodal peak values of the product were at 2.5 μm and 28.4 μm and the d₅₀ value of the product was 11.3 μm when the liquid level was 30 cm. The bimodal peak values of the product were at 2.5 μm and 11.7 μm and the d₅₀ value of the product was 6.2 μm when the liquid level was 40 cm. The bimodal values of the product were at 1.5 μm and 7.1 μm and the d₅₀ value of the product was 4.2 μm when the liquid level was 50 cm. The d₅₀ of the product decreased by 62.83% as the increased from 30 cm to 60 cm of liquid level. The corresponding small particle size peaks are more obvious, and the product particle size doublet was shifted to the left with the increase of liquid level.

The above experiments results indicate that a high liquid level would be reduced the percentage of large particles in the product and obtained smaller sized product. The possible reasons for the above phenomenon may be that particles with smaller size is lighter in weight, and under the same drag force, small particle size particles were easy to reach high liquid level positions.

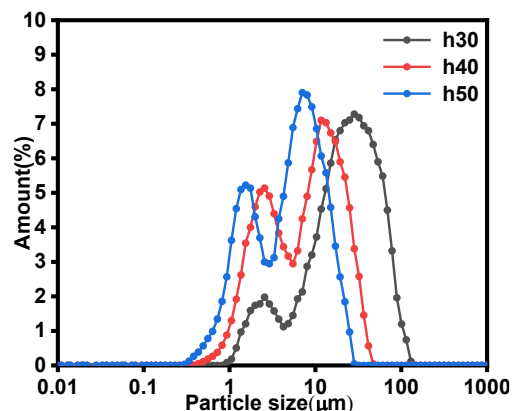


Fig.10 Effect of the liquid level on the particle size distribution. (Conditions:40 °C temperature, pH:12, 20 L·h⁻¹ CO₂ flow rate)

3.3 Characterization

3.3.1 Characterization and analysis of product

The main component of dried FGDG was CaSO₄ and the main component of the mineralization product was converted into CaCO₃ and a small amount of CaSO₄ after 60 min of mineralization reaction, as shown in Fig.11. It may be that the FGDG was deposited during the reaction, resulting in the inability of some FGDG to be in full contact with waste NaOH effluent, resulting in insufficient reaction. Fig.12a was a scanning electron microscope (SEM) image of the FGDG. The carbonization products, which was originally diamond-shaped, and the product obtained after carbonization is composed of amorphous nano-sized particles with an aggregate structure[18, 31], as shown in Fig.12b-c. EDS was used to characterize the carbonized products, as shown in Fig.12d, which mainly contained Ca, C and O and were judged to be CaCO₃ aggregates.

4 Conclusion

The effects of initial pH of NaOH effluent, CO₂ flow rate, reaction temperature, and liquid level on the mineralization efficiency of FGDG and particle size distribution of products was used to investigate by the FGDG-NaOH carbonization method. The mineralization efficiency of FGDG

increased with the increased of initial pH of NaOH effluent, but lower initial pH made it easier to obtain smaller particle products. As the CO₂ flow rate increased from 20 L·h⁻¹ to 60 L·h⁻¹, the mineralization efficiency of FGDG decreased and the particle size of the product increased. When the reaction temperature was 60 °C, the mineralization efficiency of FGDG was the highest, reaching 92%, and the particle size of the products was almost unaffected by the reaction temperature. High liquid level could be improved the mineralization efficiency of FGDG and also reduced the size of products particles. Through continuous mineralization reactions, rhombic block shaped CaSO₄ particles are transformed into aggregates of CaCO₃ nanoparticles.

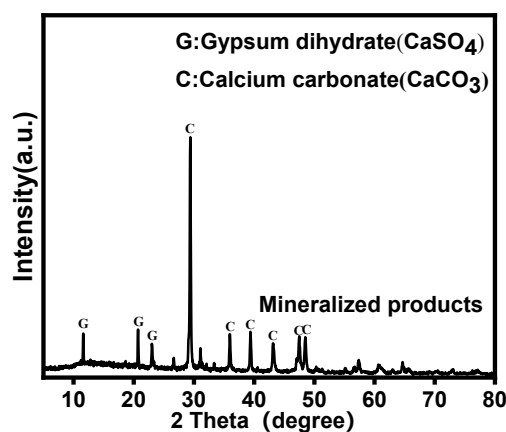


Fig.11 XRD patterns of mineralized products

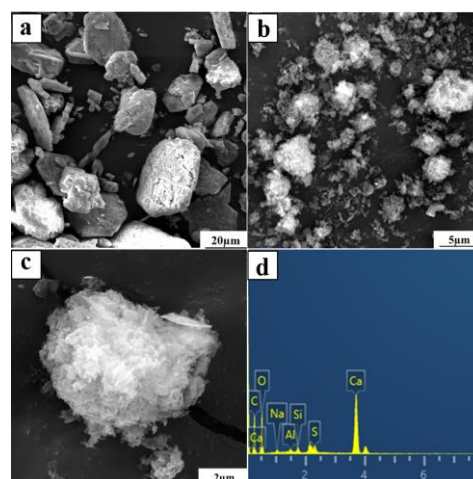


Fig.12 a: The SEM images of raw FGDG; b,-c: The SEM images of FGDG mineralization product; d: EDS mapping of FGDG mineralization product

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